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Persistent loss of biologically-rich tropical forests in the Indian Eastern Himalaya

Highlights

- We found a high rate of deforestation occurring in a state managed reserve forest in Indian Eastern Himalaya.
- Fine-scale analysis showed considerable forest loss around nesting trees for hornbills.
- Forest monitoring, protection and honest governance are required to effectively protect forests in the Eastern Himalaya biodiversity hotspot.

Abstract

Using fine-resolution satellite imagery from multiple satellite data products, we assessed the change in forest cover of a state-managed Reserve Forest (RF) located in India's Eastern Himalaya biodiversity hot-spot. 4.6% of forest cover was lost from Papum RF between 2013 and 2017 at the rate of 8.2 km^2 year^{-1}. Three species of hornbills: Great Hornbill *Buceros bicornis* Linnaeus, 1758, Wreathed Hornbill *Rhyticeros undulatus* (Shaw, 1811) and Oriental Pied Hornbill *Anthracoceros albirostris* (Shaw, 1808), that are functionally important are found here with nesting habitat in the areas affected by illegal logging. Therefore, we assessed the habitat loss within a 1 km radius around 29 nest trees. From 2011 to 2019, forest cover declined from 38.55 km^2 to 21.94 km^2 around these hornbill nest trees. Illegal logging is the main driver that is depleting forest cover within this important bird area. Our results highlight the ongoing threats to biologically-rich forests and the need for urgent measures to halt this loss. We suggest that this study has practical implications for the monitoring and governance of state-managed forests in Arunachal Pradesh.

1 Introduction

1.1 The state of forest cover in India and the Indian Eastern Himalaya

Tropical forests are not only the most biodiverse terrestrial ecosystems on Earth (Gibson et al. 2011) but also amongst the most threatened. Globally, 3.86 million square kilometres (km^2) of tree cover was lost from 2001 to 2019, with tropical forests undergoing the highest losses (Hansen et al. 2013; Global Forest Watch 2020). From 2001 to 2019 614 446 km^2 of primary forest were lost globally (Turubanova et al. 2018; Global Forest Watch 2020). Deforestation is one of the major causes of biodiversity loss across the world (Gibbs et al. 2010; Curtis et al. 2018).

India's state of forest is assessed biennially by the Forest Survey of India (FSI) and according to FSI, India has lost 80% of its native forest cover and forests continue to be lost at the rate of 1.5 to 2.7% per year. However, this does not provide an accurate estimate of the true extent of native forests and deforestation rates as these data combine native forests, secondary regrowth, plantations and cropland and do not validate classifications with ground-truthing (Puyravaud et al. 2010a,b). Puyravaud et al. (2010b) highlighted the cryptic destruction of India's native forests as a challenge to understanding the trends in the state of India's forests.

Global Forest Watch (GFW) data show that India lost about 17 955.5 km^2 of tree cover with >30% canopy cover between 2001 and 2019 amounting to 690.52 Mg of carbon dioxide (CO_2) emissions (Hansen et al. 2013; Global Forest Watch 2020). North-east India, which encompasses two global biodiversity hotspots – Indo-Burma and the Himalaya (Mittermeier et al. 2004) – appears to be severely affected by deforestation (Pandit et al. 2007). The GFW assessment estimated 12 551.4 km^2 of tree cover loss from north-east India from 2001 to 2019 (a loss of 7.8% of the tree cover extent in 2000; Global Forest Watch 2020). From 2002 to 2019, 2879 km^2 of primary forest cover was lost in north-east India (a loss of 23.72% of the total tree cover lost; Global Forest Watch 2020).

Arunachal Pradesh in north-east India is the richest terrestrial biodiversity region in India (Mishra and Datta 2007) with nearly 6000 flowering plants and 720 (54.8%) of the 1313 bird species known from India (Praveen et al. 2019). Recent research has led to the discovery of new records, range extensions and new species of fungi, plants and animals from the state (Gajurel et al. 2001; Ahti et al. 2002; Sinha et al. 2005; Athreya 2006; Tamang et al. 2008; Siliwal et al. 2015).

Forest cover in Arunachal Pradesh has been declining in the last decade although forests still cover 74% of the total land area (Global Forest Watch 2020; Table S1 and Fig. S1 in Supplementary file S1, available at https://doi.org/10.14214/sf.10373). FSI reports estimate that 486 km^2 of forest was lost from 2003 to 2017 in Arunachal Pradesh (Forest Survey of India 2003, 2005, 2009, 2011, 2013, 2017). However, GFW data show that 2093 km^2 of tree cover was lost between 2001 and 2019, comparable to a 3.3% decrease in tree cover since 2000 (Global Forest Watch 2020; Table S2 in Suppl. file S1). From 2002 to 2019, 1110 km^2 of primary forest was lost in Arunachal Pradesh comprising 55% of the total tree cover lost (Global Forest Watch 2020).

In terms of their legal status, 11.83 percent (9905.37 km^2) of the geographical area of Arunachal Pradesh is under the Protected Area (PA) network (Orchid Sanctuaries, Wildlife Sanctuaries and National Parks, some of which also include Tiger Reserves) (Singh 2017). The PAs are generally better protected than Unclassed State Forests (USF; 37.06% of area; 30 707.38 km^2) and Reserve Forests (RF; 12.04% of area; 10 089.39 km^2) attributable to stronger implementation of the country's forest and wildlife laws (Naniwadekar et al. 2015a; Singh 2017). Unclassed State Forest areas are in practice used and/or owned by the community (*de facto* rights), although recorded as being under the Forest Department. The RFs, despite being legally under the control of the state Forest Department, are often subject to various anthropogenic pressures such as agricultural expansion, conversion to plantations and/or logging (Naniwadekar et al. 2015a).

With 80% of the population practicing subsistence farming in the hilly terrain, people are primarily dependent on shifting cultivation which is mainly carried out in the USF or community forests. Shifting cultivation was estimated to cover 2040 km^2 in 2008–2009 (Wasteland Atlas 2011) but is now in decline among many communities (Teegalapalli and Datta 2016). Although shifting cultivation is usually cited as the main driver of forest loss in the state, there are several drivers of forest loss such as: agricultural expansion, growth of plantation crops such as oil palm, rubber, tea, opium, illegal logging and road expansion (Srinivasan 2014; Velho et al. 2016; Khandekar 2019). With an increasing population, need for agricultural land and development, and lack of land demarcation and cadastral surveys, there is logging (mostly in Reserve Forests) for agriculture expansion and plantations along with illegal logging in Arunachal Pradesh (Naniwadekar et al. 2015a; Velho et al. 2016; Rina 2017, 2019; Mamai 2018; Khandekar 2019).

1.2 Ethno-civil conflict and illegal logging

The main sources of revenue for Arunachal Pradesh were forest-based industries until 1996, after which the Supreme Court of India banned logging. Despite the ban, illegal clearing driven by ethno-civil conflict in Sonitpur district in neighbouring Assam resulted in the disappearance of several RFs that bordered Nameri National Park (NP) in Assam in the last two decades (Srivastava et al. 2002; Kushwaha and Hazarika 2004; Mazoomdar 2011; Velho et al. 2014; Srinivasan 2018). Srivastava et al. (2002) estimated that 232 km^2 of forests was cleared in Sonitpur District between 1994 and 2001 with the overall loss rate of 28.65%, possibly the highest deforestation rate in the country. Kushwaha and Hazarika (2004) estimated 344 km^2 forest loss between 1994 and 2002 in the Kameng and Sonitpur Elephant Reserves, while Velho et al. (2014) reported continuing forest loss in the same region around the southern boundaries of both Pakke Tiger Reserve (TR) and Nameri NP. Between 2001 and 2019, 171 km^2 of forest was lost from Sonitpur district (Global Forest Watch 2020). Forest loss over twenty-five years has resulted in substantial habitat loss for wildlife that include. the Tiger *Panthera tigris* Linnaeus, 1758, Asian Elephant *Elephas maximus* Linnaeus, 1758, and large birds such as hornbills: Great Hornbill *Buceros bicornis* Linnaeus, 1758, Wreathed Hornbill *Rhyticeros undulatus* (Shaw, 1811), and Oriental Pied Hornbill *Anthracoceros albirostris* (Shaw, 1808) (mammal taxonomic names as per Wilson and Reeder 2005; bird taxonomic names as per Integrated Taxonomic Information System 2020).

After the 1996 ban, selective logging has re-started in some forest divisions in Arunachal Pradesh since 2008–2009. However, apart from these state-controlled and permitted logging activities, ground observations and local media reports indicate that illegal logging is becoming a major driver of deforestation in Reserve Forests (Rina 2017; Sentinel Assam 2019) and other areas in Arunachal Pradesh (Press Trust of India 2017; Mamai 2018).

The Pakke TR and its surrounding RF areas are among the few remaining areas of low-elevation forest and is among the best areas for hornbills in South Asia (Datta 1998, 2001; Datta and Rawat 2003, 2004; Dasgupta and Hilaluddin 2012; Datta et al. 2012; Datta and Naniwadekar 2015) due to protection measures by forest authorities (Velho et al. 2011) and control of hunting by local people. Hornbills are an ecologically important functional group that act as effective seed dispersers (Datta 2001; Naniwadekar et al. 2015b, 2019). The main nesting habitat for the Great Hornbill, Wreathed Hornbill and Oriental Pied Hornbill, is in the low-elevation areas encompassing the Pakke TR and surrounding RFs (Datta and Rawat 2004) where illegal logging occurs. In 2012, the Hornbill Nest Adoption Programme (HNAP) was initiated to protect hornbill nest trees and nesting habitat in the Papum RF (Fig. 1) outside Pakke TR in a partnership with local communities and the state Forest Department (Datta et al. 2012; Rane and Datta 2015). Since the HNAP began, it has resulted in increased local awareness about hornbills and nest trees of three hornbill species
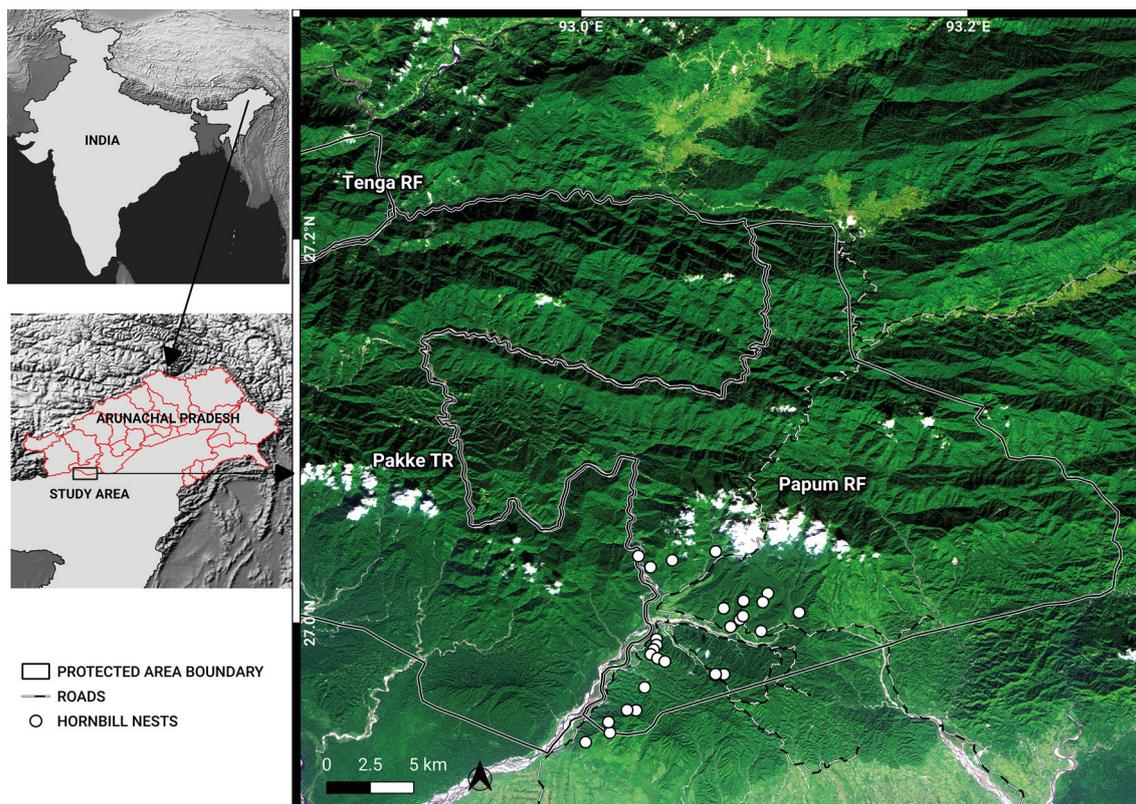
Fig. 1. A November 2018, false-colour composite image (RapidEye bands 3, 4, 1) of the study area, showing Pakke Tiger Reserve (TR), Tenga Reserve Forest (RF) and Papum RF. The administrative boundaries of Arunachal Pradesh are marked in the map of India. The border between Assam and Arunachal is also the lower boundaries of Pakke TR and Papum RF. Shades of dark green indicate forests in different elevations. Lighter green shades are cropland, secondary vegetation, flooded grasslands and bamboo. Whites are indicative of clouds, river beds and landslides. Blue depicts water. Notice the density of roads in the southwest of Papum RF.

have been protected with successful breeding and chick production. However, ground observations indicate increasing levels of illegal tree felling from 2016, with the use of chainsaws, hired labour from outside and the transport of timber outside the state.

1.3 Aims and objectives

In this study, we aimed to 1) assess the extent of forest loss in the Papum RF which adjoins the Pakke TR. Given that hornbills are a key faunal group that is functionally important and their main nesting habitat lies in the foothill areas which are affected by illegal logging, we also aimed to assess the loss of hornbill nesting habitat. Our specific purpose is to 1) estimate forest loss in the Papum RF using satellite data at a fine-scale resolution (3, 5 m) from 2013 to 2017 and 2) to determine forest loss within 1 km of hornbill nest trees at a fine-scale.

2 Methods

2.1 Study area

Papum RF covers an area of 1064 km^2 and adjoins Pakke TR (Fig. 1; 861.95 km^2, 92.5932°–93.1006°N; 26.9351°–27.2283°E). Based on historical climate data (average monthly temperature and precipitation 1970–2000; Fick and Hijmans 2017), Papum RF receives an average total annual rainfall of 2289 mm. Most (93.4%) of the rainfall occurs during the monsoon months from April to October. The dry season lasts from November to March and receives about 6.4% of the total annual rainfall. The mean (± standard deviation) maximum temperature is 25.5 °C (±1.79) during the month of August. Similarly, the mean (± standard deviation) minimum temperature is 13.6 °C (±1.87) during the month of January. The annual mean (± standard deviation) temperature is 20.6 °C (±4.35). The vegetation is classified as the Assam Valley tropical semi-evergreen forest (Champion and Seth 1968). Papum RF has a similar floral and faunal composition to the adjoining Pakke TR.

Papum RF was constituted as a RF in 1960 where all extractive activities are prohibited unless legally permitted (Indian Forest Act 1927). Part of Papum RF (346.25 km^2) is included in the buffer area of Pakke TR as per the National Tiger Conservation Authority (2012), India. Of this 318.25 km^2 is forested zone, while 28 km^2 is demarcated as multiple use area (National Tiger Conservation Authority 2012). Within Papum RF, there are 19 small towns/villages and settlements with a population of 3789 (Census of India 2011). Towards the south and east, Papum RF is bordered by Assam and Papum Pare district respectively. To the west, lies the Pakke river and Pakke TR; and to the north are the community forests of Pakke Kessang. Nameri NP in neighbouring Assam state is contiguous with Pakke TR in the south.

Although the total area is 1064 km^2, for this study, we marked out an area of 737 km^2 for classifying the forest and analysis of change in forest cover (Fig. 1). We restrict our analyses to 70% of the total area for two reasons: 1) the geographical focus of the HNAP program is within this area, 2) the boundary of entire Papum RF is uncertain and 3) the region of our analysis also forms part of the buffer area of neighboring Pakke TR. A digitized boundary of Papum RF (737 km^2, including a 500-m buffer; 92.9209°–93.2826°N; 26.9446°–27.2116°E) was used for the analyses.

2.2 RapidEye and PlanetScope satellite data processing and image classification

To conduct a supervised image classification, all satellite images were pre-processed by Planet-Labs before analysis. For example: ortho-rectified radiance/reflectance data of the RapidEye (5 m spatial resolution in 5 spectral bands) and PlanetScope (3 m spatial resolution in 4 spectral bands) constellations were obtained to ensure a complete cloud-free coverage of the Papum RF region [for a list of images analysed refer to Table S3 in Suppl. file S1; refer Planet Labs Inc. (2019) for dataset descriptions and spectral bands]. We used fine-scale satellite images for land-cover classification as this resolution can robustly resolve forest loss and other ecological phenomena below the 30-m scale (Hansen et al. 2013; Milodowski et al. 2017). Ortho-rectification (a process of image correction to account for irregular topography) is applied to ensure the same geographical region is analyzed year-to-year within a region of interest (ROI) (Tucker et al. 2004). Scenes were chosen if they were entirely cloud-free and taken by the same satellite on the same day, thereby preventing complications of image stitching and loss of information due to cloud cover. Datasets from both satellite constellations were combined to include the oldest possible year of fine-scale data (2011), and whenever RapidEye data was unavailable for analyses (example data after 2016).

Each satellite scene (or partial scene) was independently classified as forest, non-forest and logged-forest using the randomForest library 4.6-14 (Liaw and Wiener 2002) in the R software for

statistical computing (R version 3.3. R Core Development Team 2016). Ground-control polygons (GCPs) were identified within the three land-cover classes by using a combination of field sampling (using a global positioning unit) and Google Earth imagery. Forest regions comprised GCPs of closed canopy forests with little or no detectable anthropogenic disturbance. Non-forest regions comprised water bodies, grasslands, permanent settlements, sand bars and landslides. Logged-forest GCPs were defined using ground reports of active/past logging, studying satellite images at GFW deforestation hot-spots, and for roads, new clearings, plantations and fire scars. Logged-forest GCPs generally comprise areas previously under forest but currently with higher albedo than forest. The shape of the clearings are often geometrical and close to older forest clearings. Roads are linear in shape with the lower slope scarred with discarded debris. The training datasets of the above three classes consisted of at least 40 GCP's and ca. 29 million pixels, per year.

Land-cover classification of the entire Papum RF using fine-scale data was only possible for the years (2013, 2014, 2017), where these scenes fulfilled the above coverage criteria. However, the forest loss analysis around the hornbill nest trees utilized images from 2011–2019.

2.3 Land cover change around hornbill nest trees

The HNAP is confined to the lower and south-western parts of Papum RF (Fig. 1) that fall within Seijosa circle – from Darlong up to Jolly/Lanka in the north and towards the Mabuso 2/Margasso settlements to the east, within Pakke-Kessang district.

To investigate if the habitat around 29 protected hornbill nest trees (17 Oriental Pied Hornbill, 9 Great Hornbill and three Wreathed Hornbill nests) were affected by forest loss, scenes that covered >90% of the hornbill nest sites were chosen. Cloud-free, single day scenes were available and could be analysed from 2011 to 2019. This allowed us to make comprehensive fine-scale forest loss estimations for 9 years. Cloud-free satellite images for all years were from November-December, except for 2018 and 2019 which were from April–May (dry season). During the dry season, secondary vegetation in clear felled areas is visibly dissimilar from primary forest. While we do not test for this difference, we think the visible difference may be attributed to the drying and browning of vegetation in the summer season when soil moisture and rainfall are low. Secondary vegetation in winter months (post-monsoon October–February) are visibly greener as the soil moisture is still high. An identical approach (to that used for classifying forest loss in Papum RF) was implemented to classify the area around 29 hornbill nests. A 1-km buffer was created and the satellite scenes were clipped to the buffered extent (48 km^2). Three land-cover classes were defined (see above) comprising 20 GCPs and ca. 2 million pixels (RapidEye data) or ca. 5 million pixels (PlanetScope data, refer to Table S3 in Suppl. file S1).

The spatial accuracy of the land-cover classification was assessed by manual checking of the scenes combined with a stratified random sampling method (Olofsson et al. 2014). A random sample of every land-cover class in each training dataset was used to test the accuracy of the classified image providing a bias corrected estimate of land-cover area in each class. The associated standard errors, prediction accuracy and rates of commission and omission errors were estimated as recommended by Olofsson et al. (2014). The prediction accuracy and standard error of the classification (for three years' of RapidEye data) is 98.4±3.0%. For forest loss estimates around hornbill nest sites, the prediction accuracy is 96.4±7.5%. Accuracy statistics and confusion matrices for both Papum RF and the nest-sites are tabulated in the supplementary material (Table S4 and Table S5 in Suppl. file S1).

Our analyses combined results from both satellite datasets as we found the prediction accuracies to be comparable. Image classification prediction accuracies were estimated using the methods recommended in Olofsson et al. (2014). A two-sample randomization test was performed on the

distribution of all possible differences between accuracies of the observed years and then compared to the observed difference between the mean accuracies of the respective datasets (observed difference in mean accuracy = 0.03351667, p = 0.2457542; Manly 1991).

The annual rate of forest area loss was calculated on the classified land-cover images using a modified compound-interest-rate formula from Puyravaud (2003) for its mathematical clarity and biological relevance:

$$P = \frac{100}{t_2 - t_1} \ln \frac{A_2}{A_1} \tag{1}$$

where A_1 and A_2 is the forest area at time period t_1 and t_2, respectively. P is the annual percentage of area lost; ln is the natural logrithm.

3 Results

3.1 Forest loss in Papum RF: 2013–2017

There was very high forest loss in Papum RF as determined from analysis at a fine-scale resolution. Table 1 shows the loss of forest from 2013–2017 within Papum RF. While 81% of the RF was under forest in 2013, it declined to around 76% in 4 years. The area under forest, as of winter 2017, is 561 km^2 (Fig. S2 in Suppl. file S1). From 2013 to 2017, there was a loss of 32.8 km^2 of forest, with an increase in logged-forest (27.22 km^2) and of area under non-forest (4.76 km^2). Out of a total area of 737 km^2 classified, 156 km^2 was logged-forest by 2017.

Our analyses recorded forest loss to be lower in 2017 than in 2014, for two reasons: (1) an area (ca. 5 km^2) in the eastern part of Papum RF was logged in 2014 but shows growth of secondary vegetation in 2017. The spectral nature of this 5 km^2 area is very similar to forest and in 2017, the area is classified as forest. (2) Images in 2017 had a higher illumination elevation angle (46.05°), than in 2014 (39.48°), illuminating mountain slopes and forests that were previously under shadows. The illumination of river beds in 2017 also explains the increase in non-forest areas. The annual rate of forest area loss was 1.4% year^{-1} corresponding to 8.2 km^2 year^{-1}.

3.2 Forest loss around hornbill nests: 2011–2019

Forest area consistently dropped from 2011 to 2016, then increased in 2017, and decreased again up to 2019 (Fig. 2). The increase in 2017 is perhaps a measuring error due to growth of secondary vegetation in logged-forest and/or because two different satellite data products were used in 2016

Table 1. Forest loss in the Papum Reserve Forest (RF), Khellong Forest Division, Arunachal Pradesh, north-east India. Forest loss was quantified using RapidEye data for 2013, 2014, 2017. The total area of the Papum RF that was classified was 737 km^2. Numbers in parentheses indicate percentages (%).

Year	Logged-forest in km^2 (%)	Forest in km^2 (%)	Non-forest in km^2 (%)
2013	128.8 (17.5)	593.8 (80.8)	14.3 (1.9)
2014	166 (22.5)	556.5 (75.5)	14.3 (1.9)
2017	156 (21.2)	561 (76.2)	19 (2.6)
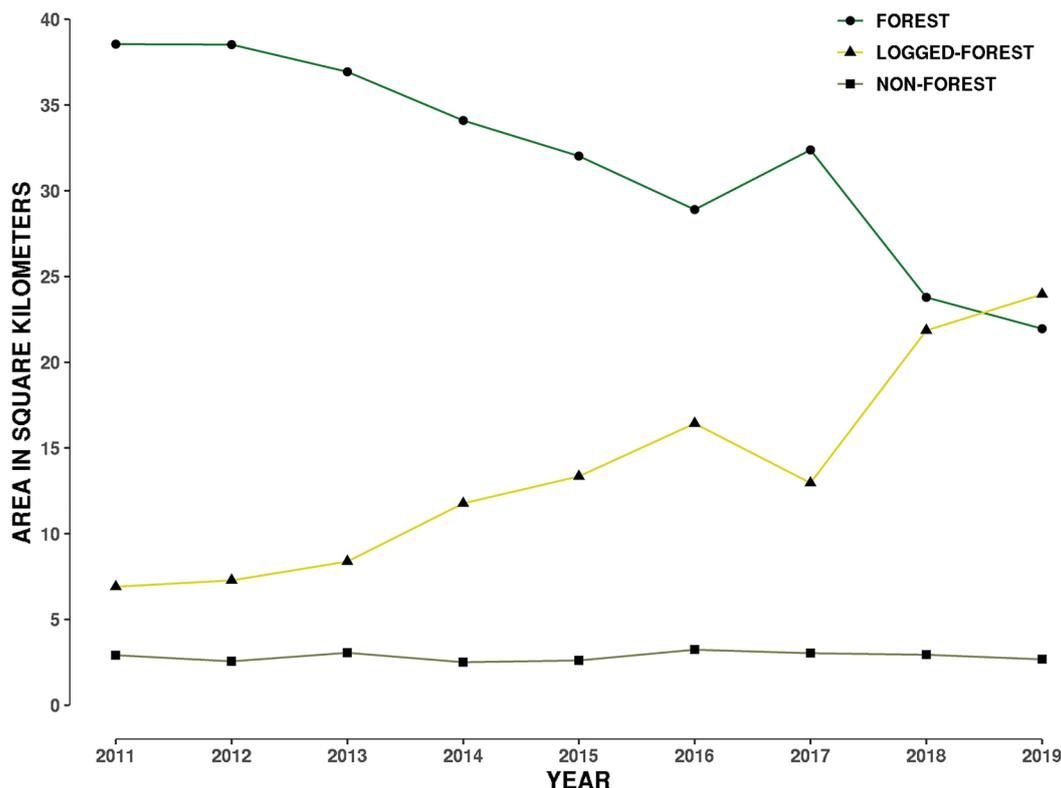
Fig. 2. Plot showing change in area of three land-cover types from 2011 to 2019 within a 1 kilometre buffer around the 29 hornbill nest sites in Papum Reserve Forest, Khellong Forest Division, Arunachal Pradesh, north-east India.

and 2017. Different satellite data products may measure vegetation spectral information differently. The error may also lie in the data collected by the land-cover class GCPs for 2017. Such an error may arise from difficulties in ascertaining secondary and primary forest in this landscape.

By 2019, only 45% of the 48 km^2 of the 1-km buffer area around 29 hornbill nests was forested as compared to 80% in 2011 (Table 2). Forest loss is also evident from the construction of roads, burn scars and clear-cut felling of primary forest areas (Fig. S3 in Suppl. file S1). During the period from 2011 to 2015, the total forest loss around nest trees was about 6 km^2, however

Table 2. Forest loss in a 1-kilometre (km) radius around 29 hornbill nests in the Papum Reserve Forest, Khellong Forest Division, Arunachal Pradesh, north-east India; 2011–2019.

Year	Forest area km^2	Percentage Forest area
2011	38.55	79.71
2012	38.52	79.66
2013	36.92	76.35
2014	34.09	70.49
2015	32.01	66.75
2016	28.90	59.50
2017	32.37	66.93
2018	23.78	48.95
2019	21.94	45.17

this increased to a loss of 4 km^2 in just one year in 2016, followed by a gain shown in 2017, with a loss of 8.59 km^2 showing up in 2018 (Table 2). In the last 9 years, there has been a total loss of 16.61 km^2 in a 1 km buffer around the 29 nest sites (Fig. 2). Annual rate of forest area loss around the nest trees was 7% year^{-1}, corresponding to 2.07 km^2 year^{-1}.

Our field observations and media reports show that tree felling increased after 2015 and coincided with the use of chainsaws and hired labour from Assam. After 2015 we frequently observed trucks transporting timber in the night using various routes for covert transport. From 2017, streams were used as roads to transport logs effectively destroying the natural aquatic ecosystem.

4. Discussion

4.1 Forest loss and its consequences

Forest loss has serious consequences for tropical biodiversity, as the destruction of suitable habitat threatens the survival of forest specialist species (Tracewski et al. 2016). Several prior studies in the area have documented the negative effects of logging on key faunal groups, vegetation structure and composition, food abundance and seed dispersal (Datta 1998; Datta and Goyal 2008; Sethi and Howe 2009; Velho et al. 2012; Naniwadekar et al. 2015b). Selective logging on a commercial scale was practiced in these RFs until the Supreme Court ban on timber extraction in 1996 (Datta 1998; Datta and Goyal 2008). However, illegal timber felling continued to occur in some pockets despite the ban. Forest loss and degradation continued due to various other factors. Several current settlements existed prior to the declaration of the RF, however the region's population has grown subsequently leading to ambiguity and conflict in terms of people's land rights and legal status of forests in the area. After devastating floods in May 2004, many families lost agricultural land to erosion, and some areas along the Assam-Arunachal border were occupied in anticipation of future needs. Over the last decade, most households stopped cultivating due to loss of land to floods and crop damage by elephants (Tiwari et al. 2017). Rubber and tea plantations also came up in the lower areas bordering Assam after 2007. These factors have led to some forest cover loss along the border areas in the 2001–2009 period. Apart from the forest loss due to these factors, until 2011–2012, timber extraction in the Seijosa area was mainly for household needs and subsistence use by people.

The loss of 32.8 km^2 of forest over 4 years within Papum RF is a cause for concern also because the area receives heavy rainfall often resulting in floods and landslides. The depletion of tropical forests in Papum RF severely threatens the future subsistence needs of the local and regional population. Although we do not explicitly test for these effects of deforestation, it is expected that landslides will increase if forest cover is lost at such a rapid rate (Bradshaw et al. 2007; Kumar and Bhagavanulu 2008; Horton et al. 2017; Stanley and Kirschbaum 2017). Soils along river valleys are destabilized accelerating river erosion rates (Horton et al. 2017) and amplifying flood risk and severity (Bradshaw et al. 2007). In mountainous regions, deforestation weakens slopes exacerbating rainfall-triggered landslides (Kumar and Bhagavanulu 2008; Stanley and Kirschbaum 2017), significantly altering river sedimentation and geomorphology (Latrubesse et al. 2009).

Additionally, deforestation alters local climate resulting in drier, warmer conditions and reduced agricultural productivity (Lawrence and Vandecar 2015), decreases access to clean drinking water (Mapulanga and Naito 2019), and has been associated with outbreaks of zoonotic diseases (Olivero et al. 2017). Furthermore, with climate change rapidly altering weather patterns, securing forests for their ecosystem services will be a pragmatic goal for all privileged and underprivileged stake-holders as per several sustainable development goals laid out by the United Nations.

Based on our study we recommend a constant monitoring of the state of forest cover by an external agency to ensure that illegal logging has been stopped. Additionally, for better governance, clarity in the use and ownership of forest land needs to be addressed under the law given that some of the designated forest area is under settlements and multiple use areas by people.

4.2 Possible effects of illegal logging on hornbills

The loss of around 35% of the forest area around the hornbill nest trees from 2011 to 2019 is alarming. From ca. 38 km^2 in 2011, the area under forest declined to 21.94 km^2 in 2019. The forest cover change analysis shows that there has been loss and degradation of the surrounding nesting habitat and hornbill food trees. Although, the HNAP has protected individual hornbill nest trees and the immediate habitat surrounding the nest tree (Rane and Datta 2015) with an estimated 138 hornbill chicks have fledged from the protected nest trees from 2012 to 2019 (Datta and Pradhan 2019), the area within 1 km radius of nest trees has been considerably degraded by illegal logging. The long-term average nest occupancy, for 8 years (2012–2019), in Pakke TR was 80%, while in Papum RF, it was 67% (±14 standard deviation). Data for the two large-bodied species (Great Hornbill and Wreathed Hornbill) was combined as there were only 3 Wreathed Hornbill nests. The long-term average nest occupancy for the large-bodied hornbills was 59.68%, while it was 72.58% for the Oriental Pied Hornbill. Nest occupancy by large-bodied hornbills was below the long-term average for these species in following years: 2013, 2015, 2016, 2017, 2018 (Fig. S4 in Suppl. file S1). These results suggest some disturbance effects on nesting success due to illegal logging. However, it is difficult to attribute it primarily to logging, as nest occupancy by hornbills can be determined by factors such as nest cavity suitability, nest occupation by other birds and mammals, inter-specific competition for nest cavities and decisions by hornbill pairs to nest in a given year.

However, the illegal logging will likely have negative consequences for hornbill nesting and persistence in the Papum RF. Tree density/basal area and food and nest tree density is considerably lower in the RF than in the Pakke TR (Datta et al. unpublished data). An earlier study has documented the negative effects of logging on hornbills and vegetation structure and composition in the area (Datta 1998; Datta and Goyal 2008). Logging also reduces food abundances for hornbills and together with hunting has consequences for seed dispersal by hornbills (Sethi and Howe 2009; Naniwadekar et al. 2015b). In any case, while most of the earlier studies have all looked at the effects of "selective" logging after some years since logging or when the logging was officially permitted before 1996, this study notes the alarming loss of forest (despite the 1996 Supreme Court ban) and the lack of any government working plan under which the current logging is occurring within Papum RF.

Hornbills are highly mobile species with large home ranges, and nesting males move from the RF to the Pakke TR to forage for fruits. (Naniwadekar et al. 2019). However, despite their ability to move between these areas, a continuing loss of forest cover will result in nest trees in the RF becoming inactive. As the forest is becoming more degraded and is being logged it has also become more common to find only nests of the more adaptable Oriental Pied Hornbill in the RF (Datta and Pradhan 2019), which is more common in open secondary forests (Datta 1998).

The tree felling occurs mainly in the drier months starting from September to March-April, but in some years, illegal logging activity has continued into the wet monsoon months. March is the beginning of the breeding season for the Great Hornbill and Wreathed Hornbill when the females start entering the nest cavities, sealing them and laying eggs. Apart from the direct loss of forest habitat and individual trees, the sound of chainsaws, movement and presence of hired labour in camps and trucks results in disturbance during this critical time in the hornbill breeding season. It is likely that hornbill breeding is being negatively affected by the ongoing illegal logging activities

which has increased in intensity in the last 2–3 years. Our long-term monitoring of hornbill roost sites located along the southern boundary of Pakke TR near the Pakke River, also shows movement of hornbills from Pakke TR to the Papum RF. The disturbance from illegal logging and loss of habitat, may also affect the use of roost sites by hornbills in the future.

4.3 Challenges and lessons for forest classification in the Eastern Himalaya

One of the challenges in our study was the strict classification of land-cover as forest, non-forest and logged-forest. Our ROI includes areas that often flood in the monsoon changing the percentages of land-cover every year. New road construction or mining in recently logged forests can be classified as non-forest, while previously cleared primary forest can show regrowth as secondary vegetation. The difficult terrain in the region makes robust collection of ground-control points challenging. Hence, we make the following suggestions: 1) dry summer season images are best to distinguish secondary and non-woody vegetation from primary forest; 2) a binary classification system of forest and non-forest; 3) forest loss estimations within a completely forested region such that loss in later years can be detected using year-to-year image subtraction techniques; and 4) pooling data from different satellite datasets can yield comparable and highly accurate estimates of land-cover change in the Eastern Himalaya.

5   Conclusions

To conclude deforestation rates are alarming in Papum RF with annual loss rates as high as 8.2 km^2 year^{-1}, as per estimates from 2013–2017 where forest cover declined to 76% of the total RF area. Our results show the loss and degradation of critical hornbill habitat underscoring the need for habitat conservation efforts to aid ongoing protection of hornbill nest trees.

We also find that both 3 and 5 m datasets have comparable land-cover prediction accuracy, indicating a trade-off between fewer bands and finer resolution versus more bands and coarser resolution imagery. Creation of land-cover maps in the Eastern Himalaya require careful interpretation of spectral data based on GCPs, as the influence of seasons and terrain on forest cover classification can be confounding. We provide suggestions that can improve efforts in monitoring and mapping forest cover in the Eastern Himalaya. Despite the limitations of our study, our work is a step towards achieving accurate forest loss estimates in an under-explored, challenging region with exceptional forests and biodiversity.

Code availability

The code for image classification is publicly available on https://github.com/monsoonforest/deforestation/blob/master/randomForest-image-classification.

Data availability

RapidEye and PlanetScope datasets are not openly available as Planet Labs (www.planet.com) is a commercial company. CS obtained the datasets through Planet Lab's Education and Research program (www.planet.com/markets/education-and-research/). The classified land-cover datasets can be made available upon request from the authors.

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Supplementary files

S1.pdf, available at https://doi.org/10.14214/sf.10373.